Activity concentrations of $^{226}$Ra, $^{232}$Th, and $^{40}$K in common maize meal consumed in Namibia and their potential radiation hazards

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Abstract
Gamma spectroscopy was performed to determine the concentrations of $^{40}$K, $^{226}$Ra and $^{232}$Th in maize meal samples collected from shop outlets and open markets in Namibia. The activity concentrations and Excess Lifetime Cancer Risk were determined using a high purity germanium (HPGe) detector. The average activity concentrations of $^{40}$K, $^{226}$Ra and $^{232}$Th were found to be $29.98 \pm 4.05$, $0.99 \pm 0.40$ and $0.35 \pm 0.08$ Bq kg$^{-1}$ in the maize meal samples. The result showed the activity concentrations of $^{40}$K significantly higher than the other radionuclides in all the maize meal samples. The average excess life cancer risk varies from $1.33 \times 10^{-13}$ to $6.05 \times 10^{-13}$ for $^{40}$K, $8.76 \times 10^{-13}$ to $1.19 \times 10^{-12}$ for $^{232}$Th and $2.43 \times 10^{-11}$ to $2.83 \times 10^{-11}$ for $^{226}$Ra. Hence, when compared with internationally acceptable limits, all values fall within the safety limits. Thus, the study concludes that the maize meals consumed in Namibia are radiologically safe for consumption.

Keywords: Natural radioactivity, maize meal, daily intake, Excess lifetime cancer risk
Introduction

The earth is continuously exposed to radiation mainly from natural and artificial sources (UNSCEAR, 2008). Natural radioactivity from uranium, thorium, and their progenies and potassium form the largest contributions to internal radiation dose received by humans owing to their wide distribution in the environment, whereas the contributions from anthropogenic sources are minimal (Al-Mastri et al., 2004). The anthropogenic release of radionuclides are localized to specific area through routine and/or accidental release and unplanned disposal of industrial and/or radioactive waste etc., that eventually find their way into the environment and food chain (Al-Mastri et al., 2004; Uwatse et al., 2015). All foods consist of naturally occurring radionuclides with concentrations varying with agricultural practices, geographical location and food type (WHO, 2011). Once radionuclides are ingested they are easily assimilated into the tissues of humans where they ultimately cause biological damages which could be for the entire lifetime (Onjefu et al., 2017).

Worldwide, maize is one among most commonly consumed staple foods. This has necessitated studies on radionuclides concentrations in maize meal samples in different regions around the globe (Nkuba and Sungita, 2017; Changizi et al., 2013; Olatunji et al., 2014). The measurement of natural radionuclides in food is relevant in quantifying radiological risk to humans due to ingestion (IAEA, 1989). This study is therefore aimed at determining the activity concentrations of natural radionuclides in different maize meal samples sold in both open markets and shop outlets in Windhoek-Namibia to determine the natural radioactivity of the maize meal samples using gamma ray spectrometry; the Hyper Pure Germanium (HPGe) detector with relative efficiency 45% of the energy resolution at 1332.5 keV energy peak of 122 keV was connected to a multi-channel analyzer (MCA). The HPGe detector was placed in the lead shield 64.80 mm thick, coated with tin and copper with a thickness of each 1 mm and 1.6 mm to reduce the effects of background radiation. Gamma ray spectrometry system was calibrated using a mixture of radioactive sources $^{60}$Co, $^{88}$Y, $^{85}$Sr, $^{57}$Co, $^{113}$Sn and $^{137}$Cs.

The concentration of radium, thorium and potassium was measured for each variety of maize meal namely: meme mahangu, namib braai pap, namib super meal (Okuryangava), maize meal, super meal, and top score. The samples were thoroughly crushed. About 500 g of each sample was taken to carry out further procedure. The samples were placed and sealed in a Marinelli beaker and left for 30 days for secular equilibrium to be attained (Onjefu et al., 2017). Initially, samples were laid in the HPGe detector, and then gamma spectrum was obtained using equation 1 (Onjefu et al., 2017):

$$C(Bq.kg^{-1}) = \frac{C_n}{\epsilon \cdot P \cdot M_s}$$

where $C$ is the activity concentration of the radionuclide in the sample in units of Bq.kg$^{-1}$, $C_n$ is the count rate of gamma rays associated with a peak per second (cps), $\epsilon$ is the detector efficiency at the specific $\gamma$-ray energy, $P_\gamma$ is the absolute transition probability of the specific $\gamma$-ray, and $M_s$ is the mass of the sample (kg).

The activity of the daughter radionuclide $^{214}$Bi and $^{214}$Pb were obtained from the 609.31 keV and 351.92 keV gamma peak respectively and were chosen as indicators of uranium ($^{238}$U) while the activity of $^{212}$Bi ($^{232}$Th) was determined from its 727.17 keV gamma ray peak and was
used as an indicator for thorium (\(^{232}\text{Th}\)).

Potassium-40 was determined by measuring the
1460.3 keV gamma rays emitted during its decay
(Jibiri and Emelue, 2008).

**Calculation of radiobiological impact**

The daily intake of radionuclides from the
activity concentrations of \(^{226}\text{Ra}\), \(^{232}\text{Th}\) and \(^{40}\text{K}\) due
to the consumption of maize meal was calculated
from equation 2 (Nahar et al., 2018).

\[
D_{\text{int}} = \frac{A_c \times A_{ig} \times D_{cf}}{Y_d}
\]  

(2)

where \(D_{\text{int}}\) represents the daily intake of radionuclides (Bq.kg\(^{-1}\)) by an individual, \(A_c\)
is the activity concentration of radionuclides (Bq.kg\(^{-1}\)), \(A_{ig}\) is the per capital per year consumption of
maize meal where for Namibia is at 44 kg
(Shifiona et al., 2016), \(Y_d\) is the days in a year.

Another radiological parameter called the
annual committed effective dose (\(\mu\text{Sv.y}^{-1}\)) to an
individual from ingestion of radionuclides in
maize meal is defined from the following formula
(Asaduzzaman et al., 2015; Nahar et al., 2018):

\[
E_{\text{eff}} = A_e \times A_{ig} \times D_{cf}
\]  

(3)

where \(E_{\text{eff}}, A_e, A_{ig}\) and \(D_{cf}\) is the annual
effective dose (\(\mu\text{Sv.y}^{-1}\)), activity concentration of
radionuclides (Bq.kg\(^{-1}\)), annual intake of maize
meal (44 kg) (kg.y\(^{-1}\)) and ingestion dose
conversion factor for radionuclides investigated
(2.8 x 10\(^{-7}\) Sv.Bq\(^{-1}\) for \(^{226}\text{Ra}\), 2.3 x 10\(^{-7}\) Sv.Bq
\(^{-1}\) for \(^{232}\text{Th}\) and 6.2 x 10\(^{-9}\) Sv.Bq\(^{-1}\) for \(^{40}\text{K}\)) (16)
(IAEA, 2011).

The gamma index \((I_\gamma)\) was calculated using
equation 4 (Senthilkumar & Narayanaswamy,
2016),

\[
I_\gamma = \frac{A_{Ra}}{150} + \frac{A_{Th}}{100} + \frac{A_{K}}{1500}
\]  

(4)

where \(I_\gamma, A_{Ra}, A_{Th}, \text{ and } A_{K}\) is the gamma index,
activity concentrations of \(^{226}\text{Ra}\), \(^{232}\text{Th}\) and \(^{40}\text{K}\)
(Bq.kg\(^{-1}\)). The total ingested dose from the
consumption of maize meal was determined using
the relationship in equation 5 (Asaduzzaman et al.,
2015):

\[
E_t = \sum (A_{fcr} \times C_{cy}) D_{cf}^f
\]  

(5)

where \(f\) represents a food group, the coefficients
\(A_{fcr}\) and \(C_{cy}\) denote the average activity
concentration of radionuclides in (Bq.kg\(^{-1}\)) and
the consumption rate per year (kg), respectively
and \(D_{cf}^f\) is the dose coefficient for an intake by
the ingestion of radioisotope, \(r\) given in unit of
(Sv.Bq\(^{-1}\)).

Excess life time cancer risk (ELCR) was
obtained from equation 6 (Nahar et al., 2018):

\[
ELCR = E_f \times D_L \times R_c
\]  

(6)

where \(ELCR\) is the life cancer risk, \(E_f\) should be
the annual effective dose \((\text{Sv yr}^{-1})\), \(D_L\) the
average lifespan (70 years) and \(R_c\) is the risk
factor \((\text{Sv yr}^{-1})\).

**Results and Discussions**

**Activity concentrations of radionuclides**

The activity concentrations of \(^{226}\text{Ra}, {^{232}\text{Th}\text{ and } {^{40}\text{K}}}\) in the maize meal samples are given in
Table 1. The activities range and mean values (in
brackets) for \(^{226}\text{Ra}, {^{232}\text{Th}\text{ and } {^{40}\text{K}}}\) are BDL - 3.20
± 1.20, (0.99 ± 0.40), BDL- 0.64 ± 0.13, (0.35 ±
0.08), and 11.04 ± 5.19 - 6.64 (29.98 ±
4.05) Bq/kg, respectively. The results show that
the mean activities of \(^{226}\text{Ra}, {^{232}\text{Th}\text{ and } {^{40}\text{K}}}\) in
the maize meal samples were below the world
average of 32, 45 and 412 Bq/kg respectively
(UNSCEAR, 2008).

The mean activity concentrations were in
the order \(^{40}\text{K} > {^{226}\text{Ra} > {^{232}\text{Th}}\). This clearly show
that \(^{40}\text{K}\) dominates over \(^{226}\text{Ra}\) and \(^{232}\text{Th}\) because
of its relative abundance in terrestrial bodies
(Wild, 1993, Dar and El-Saharty, 2013). The
meme mahangu maize meal had the highest
activity for \(^{40}\text{K}\), while namib super meal had the
highest activity of \(^{232}\text{Ra}\). The highest activity of
\(^{232}\text{Th}\) was recorded for super meal maize meal.

The differences in the activity concentrations
of \(^{40}\text{K}, {^{226}\text{Ra} and } {^{232}\text{Th}}\) in the maize meal samples
maybe attributed to differences in soil properties
and geographical setting of the soil and the type
of fertilizer added to the soil to improve soil
fertility (Karunakara et al., 2013).

The obtained mean activities of \(^{40}\text{K}, {^{226}\text{Ra} and } {^{232}\text{Th}}\) has been compared with those
reported from other countries (Table 2). It shows
that the activity of \(^{40}\text{K}, {^{226}\text{Ra} and } {^{232}\text{Th}}\) in maize
meal sold in Windhoek, Namibia is lower than the
reported maize meal range for the populations of
Tanzania, Malaysia and Iran respectively.
(Olatunji et al., 2014; Nkuba and Sungita, 2017; Changizi et al., 2013).

Table 1: Activity concentrations in maize meal samples (Bq/kg)

<table>
<thead>
<tr>
<th>Sample</th>
<th>Activity Concentrations (Bq/kg)</th>
<th>40K (Bq/kg)</th>
<th>226Ra (Bq/kg)</th>
<th>232Th (Bq/kg)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Meme Mahangu</td>
<td>50.19 ± 6.64</td>
<td>2.75 ± 1.19</td>
<td>BDL</td>
<td></td>
</tr>
<tr>
<td>Namib Braai Pap</td>
<td>24.30 ± 3.31</td>
<td>BDL</td>
<td>0.55 ± 0.13</td>
<td></td>
</tr>
<tr>
<td>Namib Super Meal</td>
<td>36.44 ± 4.86</td>
<td>3.20 ± 1.20</td>
<td>0.43 ± 0.12</td>
<td></td>
</tr>
<tr>
<td>Okuryangava Maize Meal</td>
<td>11.04 ± 1.66</td>
<td>BDL</td>
<td>0.47 ± 0.12</td>
<td></td>
</tr>
<tr>
<td>Super Meal</td>
<td>30.33 ± 4.08</td>
<td>BDL</td>
<td>0.64 ± 0.13</td>
<td></td>
</tr>
<tr>
<td>Top score</td>
<td>27.55 ± 3.72</td>
<td>BDL</td>
<td>BDL</td>
<td></td>
</tr>
<tr>
<td>Minimum</td>
<td>11.04 ± 1.66</td>
<td>BDL</td>
<td>BDL</td>
<td></td>
</tr>
<tr>
<td>Maximum</td>
<td>50.19 ± 6.64</td>
<td>3.20 ± 1.20</td>
<td>0.64 ± 0.13</td>
<td></td>
</tr>
<tr>
<td>Mean</td>
<td>29.98 ± 4.05</td>
<td>0.99 ± 0.40</td>
<td>0.35 ± 0.08</td>
<td></td>
</tr>
</tbody>
</table>

BDL= Below Detection Limit

Table 2: Average activity concentrations (Bq.kg⁻¹) of ⁴⁰K, ²²⁶Ra and ²³²Th in maize meal samples investigated in this study, compared with those reported from other countries.

<table>
<thead>
<tr>
<th>Origin</th>
<th>⁴⁰K</th>
<th>²²⁶Ra</th>
<th>²³²Th</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>Namibia</td>
<td>29.98</td>
<td>0.99</td>
<td>0.35</td>
<td>Present study</td>
</tr>
<tr>
<td>Tanzania</td>
<td>42.0-434.6</td>
<td>-</td>
<td>2.2-38.9</td>
<td>Nkuba and Sungita (2017)</td>
</tr>
<tr>
<td>Iran</td>
<td>91.73</td>
<td>1.67</td>
<td>0.5</td>
<td>Changizi et al., (2013)</td>
</tr>
<tr>
<td>Malaysia</td>
<td>26.4-129</td>
<td>0.05-19.18</td>
<td>ND</td>
<td>Olatunji et al., (2014)</td>
</tr>
</tbody>
</table>

ND = No Data

Table 3: Daily intake (Din) of ⁴⁰K, ²²⁶Ra and ²³²Th and annual effective dose due to the consumption of maize meal.

<table>
<thead>
<tr>
<th>Sample ID</th>
<th>⁴⁰K (Bq.d⁻¹)</th>
<th>²²⁶Ra (Bq.d⁻¹)</th>
<th>²³²Th (Bq.d⁻¹)</th>
<th>Annual effective dose, E_eff (μSv.y⁻¹)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Meme Mahangu</td>
<td>6.05</td>
<td>0.33</td>
<td>-</td>
<td>1.37 x 10⁻⁰⁶</td>
</tr>
<tr>
<td>Namib Braai Pap</td>
<td>2.93</td>
<td>-</td>
<td>0.07</td>
<td>6.63 x 10⁻⁰⁶</td>
</tr>
<tr>
<td>Namib Super Meal</td>
<td>4.39</td>
<td>0.39</td>
<td>0.05</td>
<td>9.94 x 10⁻⁰⁶</td>
</tr>
<tr>
<td>Okuryangava Maize Meal</td>
<td>1.33</td>
<td>-</td>
<td>0.06</td>
<td>3.01 x 10⁻⁰⁶</td>
</tr>
<tr>
<td>Super Meal</td>
<td>3.66</td>
<td>-</td>
<td>0.08</td>
<td>8.27 x 10⁻⁰⁶</td>
</tr>
<tr>
<td>Top score</td>
<td>3.32</td>
<td>-</td>
<td>-</td>
<td>7.52 x 10⁻⁰⁶</td>
</tr>
<tr>
<td>Average Value</td>
<td>3.61</td>
<td>0.36 ± 0.03</td>
<td>0.07</td>
<td>8.17 x 10⁻⁰⁶</td>
</tr>
</tbody>
</table>

ND = No Data

Daily intake and Annual Effective Dose

The daily intake of ⁴⁰K, ²²⁶Ra and ²³²Th and annual effective dose due to the consumption of maize meal was calculated and the results shown in Table 3. The estimated daily intakes of activity concentrations of ⁴⁰K, ²²⁶Ra and ²³²Th into the human body by ingestion of the maize meal varied from 1.33 to 6.05 Bq.d⁻¹ with an average value of 3.61±0.79 for ⁴⁰K, 0.33 to 0.39 Bq.d⁻¹ with an average value of 0.36±0.03 for ²²⁶Ra, and 0.05 to 0.08 Bq.d⁻¹ with an average value of 0.07±0.008 for ²³²Th, respectively. The daily intake of ⁴⁰K was found to be significantly higher than the other radionuclides in all the maize meal samples. This
radionuclide is an essential element for the human body. However, the use of fertilizers for improve crop yield may have informed the high levels of $^{40}$K activity concentrations in all the samples (Jayasinghe et al., 2020). The annual effective dose in µSv.y$^{-1}$ due to ingestion of maize meal were in the range $3.01 \times 10^{-06}$ to $1.37 \times 10^{-05}$ for $^{40}$K, $3.38 \times 10^{-05}$ to $3.95 \times 10^{-05}$ for $^{226}$Ra and 6.45 x $10^{-06}$ for $^{232}$Th respectively. The Annual effective dose for Namib Super Meal was the highest for all radionuclides except for $^{232}$Th. These values are however far below the threshold dose values for safety from ingestion of uranium (6.3 µSv.y$^{-1}$) and thorium (0.38 µSv.y$^{-1}$) radionuclides as reported by UNSCEAR (2000).

**Gamma index and excess lifetime cancer risk**

Table 4 present the gamma index and excess lifetime cancer risk. The calculated values of the gamma Index ranges from 0.01 (Okuryangava Maize Meal) to 0.03 (Super Meal), with a mean value of 0.03±6.66. The $I_{\gamma}$ values for all maize meal sample were lower than the critical value of unity (UNSCEAR, 2000). The Excess life time cancer risk (ELCR) varied from $1.33 \times 10^{-13}$ to $6.05 \times 10^{-13}$ for $^{40}$K, $2.43 \times 10^{-11}$ to $2.83 \times 10^{-11}$ for $^{226}$Ra and $7.99 \times 10^{-13}$ to $1.02 \times 10^{-12}$ which are lower than the acceptable ELCR limit as set by USEPA (1989). The United States Environmental Protection Agency (USEPA) considers acceptable for regulatory purposes an excess life time cancer risk of between $1 \times 10^{-6}$ and $1 \times 10^{-4}$ (USEPA, 1989).

**Conclusions**

Concentration of natural radionuclides in various types of maize meal samples consumed in Namibia has been determined. The mean activity concentrations of $^{40}$K, $^{226}$Ra and $^{232}$Th were found to be 29.98 ± 4.05, 0.99 ± 0.40 and 0.35 ± 0.08 Bq.kg$^{-1}$ in maize meal samples. The daily intake of radionuclides was found to be in the order $^{40}$K $\geq$ $^{226}$Ra $\geq$ $^{232}$Th. From the measured values of the activity concentrations, the effective dose, gamma index and excess lifetime cancer risk were calculated and found to be lower than internationally accepted threshold limits. The study reveals that radionuclides intake from consumption of maize meals pose no radiological threat to members of the public in Windhoek-Namibia. The findings from this study will help in establishing baseline information for radionuclides exposure to members of the public due to consumption of maize meal.

**Table 5: Gamma index, alpha index and excess life time cancer risk**

<table>
<thead>
<tr>
<th>Sample ID</th>
<th>Gramma Index ($I_{\gamma}$)</th>
<th>$^{40}$K (Bq.kg$^{-1}$)</th>
<th>$^{226}$Ra (Bq.kg$^{-1}$)</th>
<th>$^{232}$Th (Bq.kg$^{-1}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Meme Mahangu</td>
<td>0.05</td>
<td>6.05 x 10^{-13}</td>
<td>2.43 x 10^{-11}</td>
<td>-</td>
</tr>
<tr>
<td>Namib Braai Pap</td>
<td>0.02</td>
<td>2.93 x 10^{-13}</td>
<td>-</td>
<td>1.02 x 10^{-12}</td>
</tr>
<tr>
<td>Namib Super Meal</td>
<td>0.05</td>
<td>4.39 x 10^{-13}</td>
<td>2.83 x 10^{-11}</td>
<td>7.99 x 10^{-13}</td>
</tr>
<tr>
<td>Okuryangava Meal</td>
<td>0.01</td>
<td>1.33 x 10^{-13}</td>
<td>-</td>
<td>8.76 x 10^{-13}</td>
</tr>
<tr>
<td>Super Meal</td>
<td>0.03</td>
<td>3.65 x 10^{-13}</td>
<td>-</td>
<td>1.19 x 10^{-12}</td>
</tr>
<tr>
<td>Top score</td>
<td>0.02</td>
<td>3.32 x 10^{-13}</td>
<td>-</td>
<td>-</td>
</tr>
<tr>
<td>Mean</td>
<td>0.03±6.66</td>
<td>3.61 x 10^{-13}</td>
<td>8.77 x 10^{-12}</td>
<td>6.48 x 10^{-13}</td>
</tr>
</tbody>
</table>

**References**


Asaduzzaman, K.H., Khandaker, M.U., Amin, Y. M., Bradley, D.A., Mahat, R. H. and Nor,


Karunakara N., Rao, C, Ujwal, P, Yashodhara, I., Kumara, S., and Ravi, P.M. (2013). Soil to rice transfer factors for $^{226}$Ra, $^{228}$Ra, $^{210}$Pb, $^{40}$K and $^{137}$Cs: a study on rice grown in India, *Journal of Environmental Radioactivity 118*, 80-92


